



Are *Calanus* spp. shifting poleward in the North Atlantic? A habitat modelling approach

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In the last decade, the analysis based on Continuous Plankton Recorder survey in the eastern North Atlantic Ocean detected one of the most striking examples of marine poleward migration related to sea warming. The main objective of this study is to verify the poleward shift of zooplankton species (*Calanus finmarchicus*, *C. glacialis*, *C. helgolandicus*, *C. hyperboreus*) for which distributional changes have been recorded in the North Atlantic Ocean and to assess how much of this shift was triggered by sea warming, using Generalized Additive Models. To this end, the population gravity centre of observed data was compared with that of a series of simulation experiments: (i) a model using only climate factors (i.e. niche-based model) to simulate species habitat suitability, (ii) a model using only temporal and spatial terms to reconstruct the population distribution, and (iii) a model using both factors combined, using a subset of observations as independent dataset for validation. Our findings show that only *C. finmarchicus* had a consistent poleward shift, triggered by sea warming, estimated in 8.1 km per decade in the North Atlantic (16.5 per decade for the northeast), which is substantially lower than previous works at the assemblage level and restricted to the Northeast Atlantic. On the contrary, *C. helgolandicus* is expanding in all directions, although its northern distribution limit in the North Sea has shifted northward. *Calanus glacialis* and *C. hyperboreus*, which have the geographic centres of populations mainly in the NW Atlantic, showed a slight southward shift, probably responding to cool water penetrating southward in the Labrador Current. Our approach, supported by high model accuracy, shows its power in detecting species latitudinal shifts and identifying its causes, since the trend of occurrence observed data is influenced by the sampling frequency, which has progressively concentrated to lower latitudes with time.

Keywords: *Calanus*, climate change, habitat model.

Introduction

Data from Continuous Plankton Recorder (CPR) survey in the Northeast Atlantic Ocean indicate that zooplankton exhibit distribution range shifts in response to global warming that are among the fastest and largest of any marine or terrestrial group (Beaugrand *et al.*, 2002; Lindley and Daykin, 2005; reviewed by Richardson, 2008). Habitat models based on CPR data have also been developed to provide projections of future climate-driven shifts (Helaouet and Beaugrand, 2007, 2009; Beaugrand *et al.*, 2008; Helaouet *et al.*, 2011; Reygondeau and Beaugrand, 2011). Factors that control the spatial distribution of populations can be grouped into two main categories, external and internal (Planque

et al., 2011). External controls, which are often referred to as environmental controls, are independent of population state, and they are the basis of the ecological niche theory (Hutchinson, 1957). Different statistical and mathematical techniques have been developed to model the spatial distribution of species, also termed species distribution models (Elith and Leathwick, 2009), climate envelope models (Pearson and Dawson, 2003), habitat distribution models (Guisan and Zimmermann, 2000), or habitat suitability models (Hirzel *et al.*, 2002). A climate envelope model generally characterizes a set of suitable habitats for a species derived from their present geographic location. As, in general, habitat models are constructed from the associations between the extant geographic

position of a species' occurrence and its climate, there is difficulty in testing these models under different climates (Lawing and Polly, 2011).

The second type of factors affecting the spatial distribution of populations is the internal control, which are directly linked to the population prospects (reproduction, mortality and migration rate, that for copepods is mainly based on advection). These include density-dependent processes, the effect of demographic structure, and biogeographical processes (Ohman and Hirche, 2001; Planque et al., 2011). These two factor types have traditionally differentiated both the conceptual view of species control (niche theory vs. population dynamics) and the modelling approaches (habitat models vs. individual-based models; Runge et al., 2005). When modelling the pelagic habitat of plankton, dispersal limitation is often neglected, although it can play an important role on community structure (Irigoiien et al., 2011; Chust et al., 2013). Research efforts need to focus on integrating the two mechanisms in the modelling approach for zooplankton species, as attempted for fish and invertebrates (Cheung et al., 2009).

Time-series of spatially explicit biological data such as zooplankton occurrence are challenging to analyse because of non-uniform sampling across time and space. Very few predictive models have been validated using independent data (but see Lewis et al., 2006; Llope et al., 2012). Past works showed shift of populations (zooplankton assemblages) based on northward/southward limits of a restricted region (Northeast Atlantic) and did not encompass the overall distribution range of species (e.g. Beaugrand et al., 2002). Therefore, data reconstruction methods and models encompassing all North Atlantic and including both niche and population factors are needed to avoid bias in trend analysis and to explore possible causes of population shifts. The application of habitat models to case species also requires key steps to be accurately undertaken to avoid model overfitting (Burnham and Anderson, 2002), such as the selection of the explanatory variables and model validation. Model validation is one of the critical steps to extrapolate models to conditions outside of those used to generate the model, such as to other regions (Valle et al., 2011) and to future climate (Lawing and Polly, 2011).

Our objective is to develop models of habitat suitability (at species level) and data reconstruction for analysing past changes in zooplankton species (e.g. *Calanus* spp.) due to prominent climate drivers, particularly sea warming within the North Atlantic (Figure 1), using Generalized Additive Models (GAMs; Hastie and Tibshirani, 1990). Zooplankton are good indicators of climate change for several reasons (Richardson, 2008): highly sensitive to temperature, most species are short-lived so there can be tight coupling of climate and population dynamics, copepods are usually not commercially exploited avoiding confusion with trends in exploitation, and the distribution of zooplankton can accurately reflect temperature and ocean currents because plankton are free floating. In this work, habitat suitability models were built for each one of the four copepod species (*Calanus finmarchicus*, *C. glacialis*, *C. helgolandicus*, and *C. hyperboreus*) within the temperate to Subarctic North Atlantic from 1959 to 2004. These species, especially *C. finmarchicus*, dominate zooplankton biomass in N. Atlantic; as the one of the main grazers they transfer energy from phytoplankton to higher trophic levels such as commercially exploited fish (Planque and Batten, 2000; Richardson, 2008). The aim of this paper is to test whether population gravity centres of key taxa such as *Calanus* spp. shift poleward, and if so to explore whether this movement could be associated with sea warming. To this end, we compare time-series of observed data

with a series of GAMs as numerical experiments: (i) a model using only temporal and spatial terms, as a data reconstruction method, to simulate population movement, (ii) a model using only climate factors (i.e. niche-based model) to simulate the habitat suitability, and (iii) a model using both climate and spatio-temporal factors combined. Subsequently, for all the models, the distribution of the species was predicted over the whole spatio-temporal domain. Thus, the rate of change of the geographic gravity centre of the modelled occurrence probability allowed us studying the shifts of the species population and habitat suitability along time.

Material and methods

Biological data and sampling frequency

Data on the abundance [mean density (ind. m⁻³) of four species (*C. finmarchicus*, *C. glacialis*, *C. helgolandicus*, *C. hyperboreus*), total diatoms and dinoflagellates, and the phytoplankton colour index (PCI) were obtained from the CPR database. The CPR survey is an upper layer plankton monitoring programme that has regularly collected samples, at monthly intervals, in the North Atlantic and adjacent seas since 1946 (Warner and Hays, 1994). Water enters the CPR through a small aperture at the front of the sampler and travels down a tunnel where it passes through a silk filtering mesh of 270 µm before exiting at the back of the CPR. The plankton filtered on the silk is fixed to a final concentration of 4% formalin into a tank within the CPR body. On return to the laboratory, the silk roll is unwound and cut into sections corresponding to 10 nautical miles (~3 m³ of seawater filtered), the greenness of the silk is assessed and the plankton microscopically identified (Jonas et al., 2004). The CPR data used in the present study represent monthly data collected between 1959 and 2004 within 35 to 65°N and 75°W to 9°E. Data were gridded using the inverse-distance interpolation method (Isaaks and Srivastava, 1989), in which the interpolated values were the nodes of a 1° by 1° grid. Total diatoms and dinoflagellates represent an aggregation of data from 66 and 39 genera, respectively. *Calanus glacialis* and *C. hyperboreus* are Arctic species, whereas *C. finmarchicus* is a Subarctic species and *C. helgolandicus* dwells on temperate regions (Beaugrand et al., 2002). *Calanus finmarchicus* overlaps in size range with *C. helgolandicus* and *C. glacialis* which is the largest among them. The *C. finmarchicus*, *C. helgolandicus*, *C. glacialis*, and *C. hyperboreus* data here used are represented by the fifth copepodite and adult stages (i.e. CV-CVI). We use the PCI, which is a visual assessment of the greenness of the silk, as an indicator of the distribution of total phytoplankton biomass across the Atlantic basin (Batten et al., 2003; Richardson et al., 2006). This index covaries positively with both fluorimeter and satellite measures of chlorophyll (Batten et al., 2003; Raitos et al., 2005).

A non-uniform distribution of sampling in space and time might influence the subsequent analysis of the poleward shift of species population gravity centres. Therefore, sampling frequency have been analysed as a function of year, latitude, and longitude to test the uniformity of sampling effort over the period.

Environmental data and climate variability

Environmental data compiled had a spatial resolution of 1° longitude and 1° latitude and cover the entire domain defined. Sea surface temperature (SST), salinity, vertical velocity, and sea level anomaly were extracted from the reanalysis OS3 ECMWF (European Centre for Medium-Range Weather Forecasts) model and downloaded from CliSAP-Integrated Climate Data Center

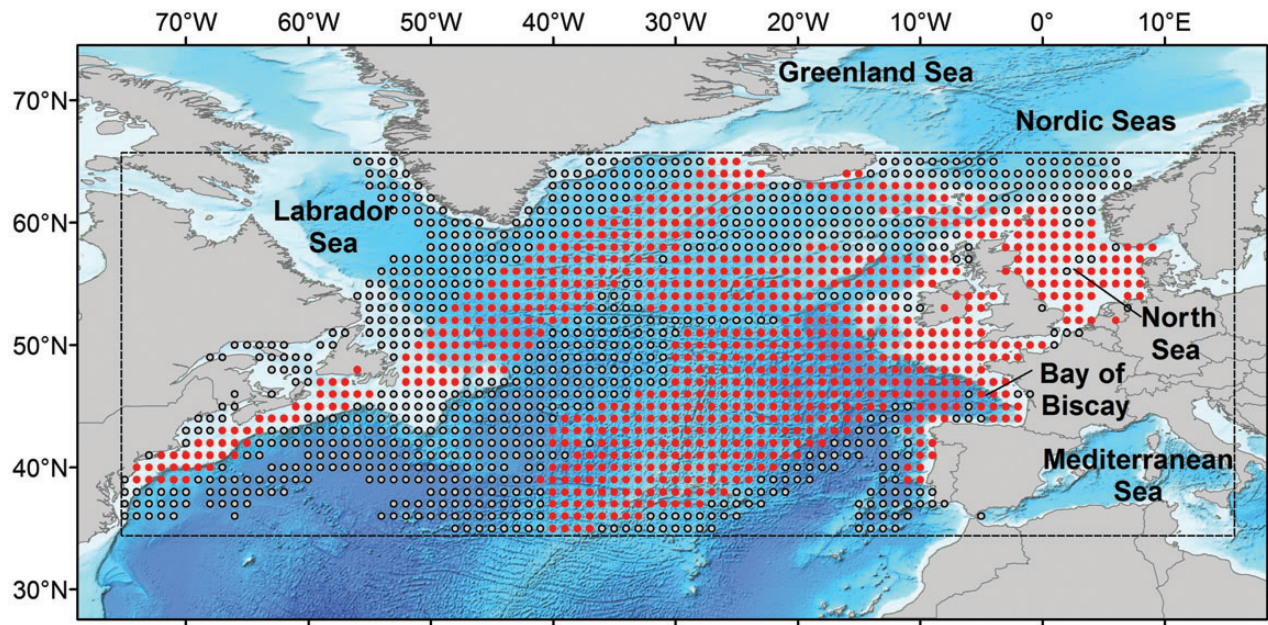


Figure 1. The North Atlantic basin. The domain of the studied area is 35° to 65°N and 75°W to 15°E. Source of Bathymetry: ETOPO1, NOAA, Amante and Eakins (2009). Transparent dots are those sampled in the overall period (1959–2004). Red dots are those sampled in 2004.

(ICDC). Bathymetry was extracted from the ETOPO1 global model (NOAA).

The time-series trends of SST were analysed over the period 1959–2004 to test whether isothermals increased at all latitudes. In addition, the overall mean SST was analysed over the period and map differences between cold and warm periods were calculated.

Habitat modelling and occurrence reconstruction

The approach consisted in modelling the species occurrence as a function of environmental factors (SST and salinity, vertical velocity), surrogates of environmental factors and population features (latitude and longitude, month), temporal trend (year), and potential food resources (PCI, total abundance of diatoms and of dinoflagellates), using GAMs (Hastie and Tibshirani, 1990). GAMs enable the fit of non-linear models for a wide family of statistical distributions. The presence/absence of each *Calanus* spp. was assumed to follow a binomial error distribution and the logit was considered as the link function. Overfitting was prevented by restricting the degrees of smoothness as explained below and by visualizing the response for each variable. The CPR dataset encompasses 178 910 samples across the spatial domain and irregularly distributed at year and monthly basis (see 2004 data as an example in Figure 1).

SST and salinity were used because of their recognized influence, either direct or indirect, on the spatial distribution of *Calanus* spp. (e.g. Helaouet and Beaugrand, 2007). Vertical velocity was used as a surrogate of upwellings. For environmental variables, the degree of smoothness of model terms was restricted from 3 to 5 to assume a unimodal niche model (*sensu* Hutchinson, 1957), but allowing asymmetry since interactions between species and extreme environmental stress may cause skewed responses (Oksanen and Minchin, 2002). In this sense, GAMs provide a more realistic solution than rectilinear climate envelope models or ellipsoidal climate envelope GLMs (Oksanen and Minchin, 2002). The CPR phytoplankton indices (i.e. PCI, diatoms, and dinoflagellates abundance) were

tested individually as phytoplankton is a key food resource for *Calanus* spp. (Irigoin *et al.*, 1998); however, because the data relative to these indices were not always available to cover the spatial and temporal domain as the other environmental variables, they were not used for spatial and temporal reconstruction of the habitat suitability model and they were not included in the overall model.

Month and geographic variables (i.e. latitude and longitude) are considered here as surrogates of population features or environmental factors not considered (Legendre, 1993). For instance, spatial terms might be proxies of overwintering areas, since close vicinity to an overwintering site has been suggested to be the main prerequisite for a region or water mass to hold a large population of *C. finmarchicus* (e.g. Head *et al.*, 2000; Speirs *et al.*, 2004; Torgersen and Huse, 2005). Spatial terms (i.e. latitude and longitude) were considered with interaction and using a bivariate smooth function. Month information was considered as a categorical variable; thus, a parametric coefficient was estimated for each month. To analyse the temporal trend, year was considered as explanatory variable and its GAM smoothness term was restricted to 5 degrees of freedom.

We built and compared three different models for each species. (i) A spatial and temporal model. A GAM using spatial (latitude, longitude) and temporal (month, year) terms only. This model is employed as data reconstruction method, thus, to identify trends on population over the analysed period. (ii) An SST model. A GAM based on only SST. As this model is driven by SST trend, it serves to identify species habitat suitability shifts. The comparison of the output of this model with the previous one permits to verify whether populations have shifted due to sea warming or if only their habitat suitability has shifted. (iii) A combined model. A GAM based on spatial, temporal, and environmental factors. As this model incorporates all factors, it has the potential to be the most accurate in predicting overall habitat suitability among the three model types, and it is used also to assess the relative contribution of spatio-temporal factors with respect to the environment.

Several steps had been undertaken: first, we have built a GAM for each variable independently. Second, we have selected the best model by removing variables that are not statistically significant or explaining the deviance of species occurrence by $< 1\%$. The explained deviance, $1 - (\text{residual deviance})/(\text{null deviance})$, is the equivalent to R^2 in least-squares models (Guisan and Zimmermann, 2000). Third, we have improved the model using a forward stepwise procedure, i.e. by adding variables according to the explained deviance, and removing those that do not improve significantly the model or $< 1\%$.

Model validation

We validated the models based on the cross-validation resampling procedure, which use independent datasets for model building and model validation (Burnham and Anderson, 2002). The comparison between the accuracy of the model (that using all observations to build the model) and that of cross-validated permits the detection of model overfitting, which highly reduce the use of such models for extrapolation. In particular, we used k -fold partitioning of a dataset (Hijmans et al., 2012), where each record in the dataset is randomly assigned to a group. In our case, group numbers are between 1 and k , with $k = 5$, hence, 80% of the CPR observations are used for model building, and the remaining 20% (i.e. independent) observations are used for model validation in an iterative procedure that is repeated five times.

The accuracy of the model and the five replicate model cross-validations have been evaluated using area under the receiver operating characteristic curve (AUC; Fielding and Bell, 1997; Raes and ter Steege, 2007) and accuracy indices derived from the confusion matrix (VanDerWal et al., 2012). To this end, first, the modelled probability of species presence was converted to either presence or absence using probability thresholds obtained using two criteria: sensitivity (true predicted presences) is equal to specificity (true predicted absences), and maximization of sensitivity plus specificity, following Jimenez-Valverde and Lobo (2007). Given the defined threshold value, a confusion matrix (also called an error matrix, Congalton, 1991), which represents a cross-tabulation of the modelled occurrence (presence/absence) against the observations data, was calculated. An overall accuracy measure was computed from the confusion matrix which is the proportion of the presence and absence records correctly identified. Both overall accuracy and AUC values range between 0.5 (random sorting) and 1 (perfect discrimination).

Time-series analysis of the geographic gravity centre

For all three model types, the probability of species occurrence was predicted over the whole spatial and temporal domain. Subsequently, we computed the geographic gravity centre of the predicted maps of probability of the presence of each species from January 1959 to December 2004. The gravity centre is defined as the mean geographic location of a population (Wuillez et al., 2009). The changes in the gravity centre's coordinates (longitude and latitude) were used to describe the changes in the geographic distribution of the habitat suitability in the *Calanus* spp. along time. As a way to test the sensitivity of this approach to detect geographic shifts triggered by environmental change, first, this statistic was calculated from the habitat suitability models for each time-step (i.e. a month) and subsequently represented along the seasonal cycle. Latitudinal trends estimated from these three models were compared with observed data (both abundance and presence/absence data) and sampling frequency.

RESULTS

Climate variability and change

The analysis of time-series trends over the period 1959–2004 indicated that SST at all latitudes analysed increased between 0.0240 and $0.0088^\circ\text{C year}^{-1}$ ($p < 0.0001$, $p = 0.018$, respectively). Specifically, the warming period started mainly at circa 1970 with a rate of increase of $0.028^\circ\text{C year}^{-1}$ ($p < 0.0001$) as showed by the mean SST (Figure 2a). The spatial difference of SST between cold (1969–1979) and warm (1998–2004) periods (Figure 2b) showed that warming in the NW Atlantic was more heterogeneous than in the NE Atlantic with three hotspots of warming and a region of slight cooling.

CPR data sampling frequency

Sampling frequency has a strong variability throughout years, with maximum effort in the 1960s and minimum in the 1980s (Figure 3a). The mean sampling frequency at the beginning of the period is more concentrated in higher latitudes and decreases 3° on average over the period, with moderate variability (Figure 3b). The mean sampling frequency is more concentrated in the eastern part of the North Atlantic during 1959–1961 and during 1987–1989 and in the western part in the remaining period (Figure 3b). There is a non-linear trend in the sampling frequency for the geographic longitude along the period.

Habitat models

First, each term (environmental, spatial, or temporal) was separately tested for each species using the GAM (Table 1). This indicates that the PCI and the total abundance of diatoms and dinoflagellates

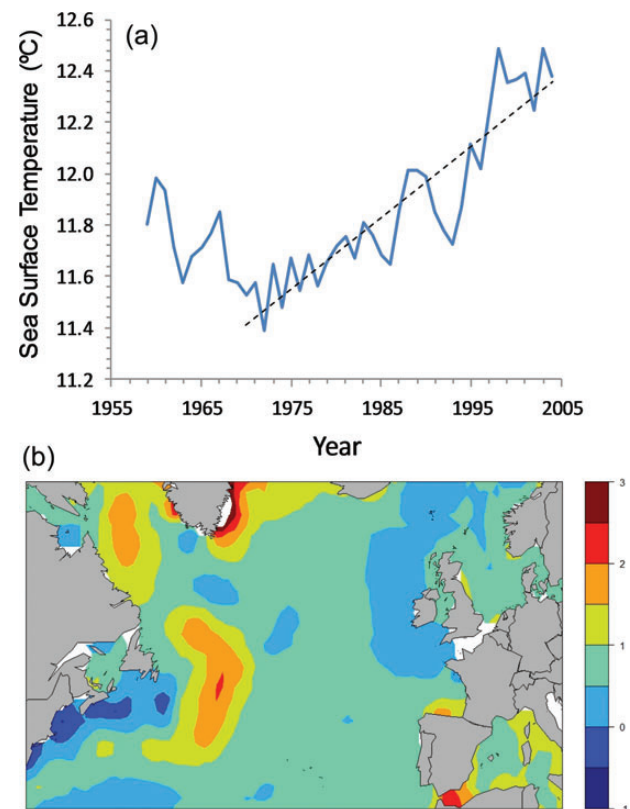


Figure 2. (a) Mean SST trend; (b) difference of SST between 1969–1979 (cold period) and 1998–2004 (warm period).

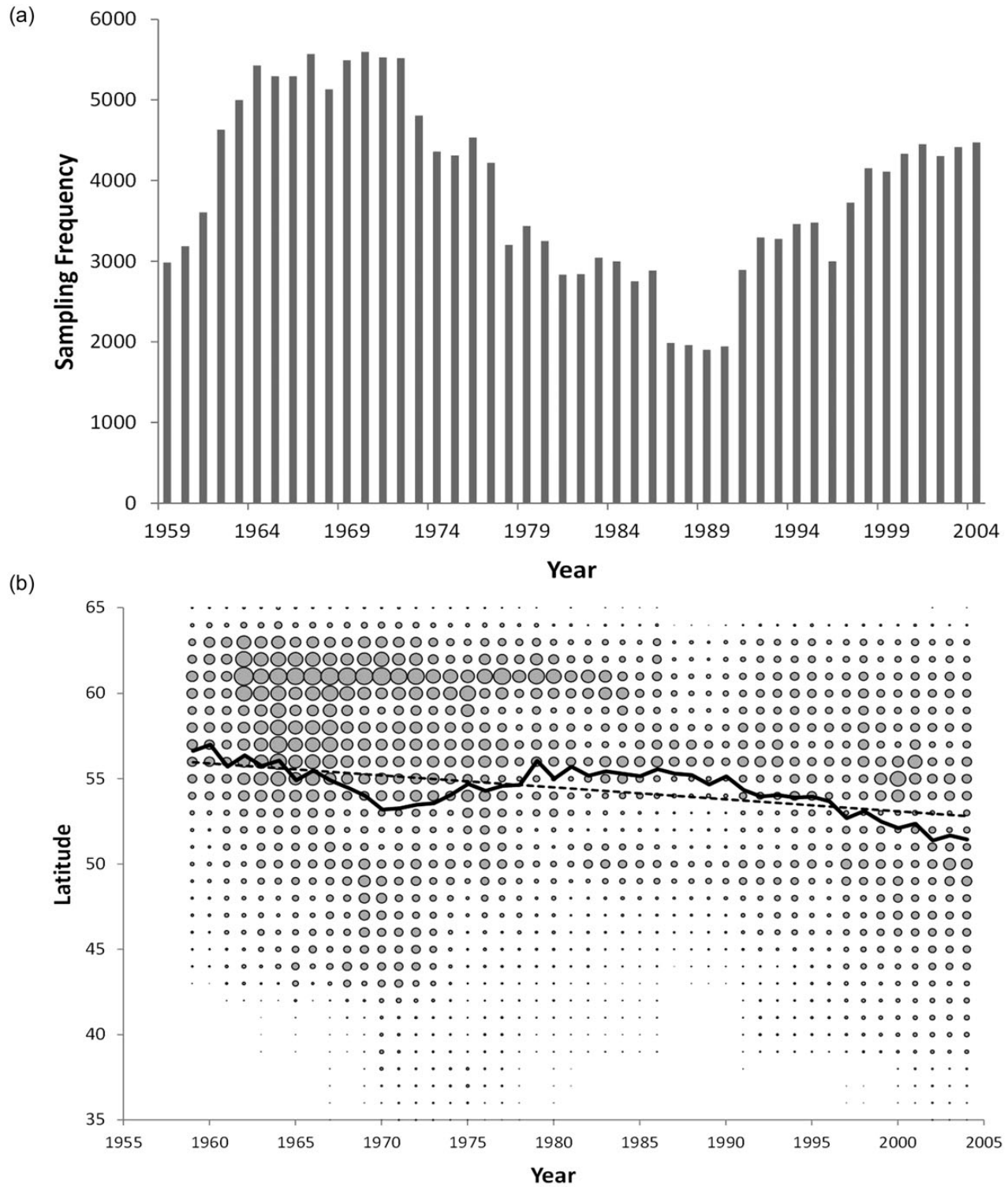


Figure 3. Sampling frequency as a function of year (a), as a function of year and latitude (b), and as a function of year and longitude (c). In (b) and (c), the size of the circle is proportional to the sampling frequency, the black line indicates the mean value at the corresponding year, and the dotted line indicates the linear temporal trend.

accounted for a limited deviance of species occurrence. Subsequently, the three model types were fitted to each species according to the methodology described and the explained deviance compared. In all *Calanus* spp., spatial and temporal models accounted for

33–45% of species occurrence, whereas environmental variables accounted for 17–31% (Supplementary S11). Specifically, the SST model explained between 26 and 10% of the variation according to species. Environmental factors contributed very little to the combined

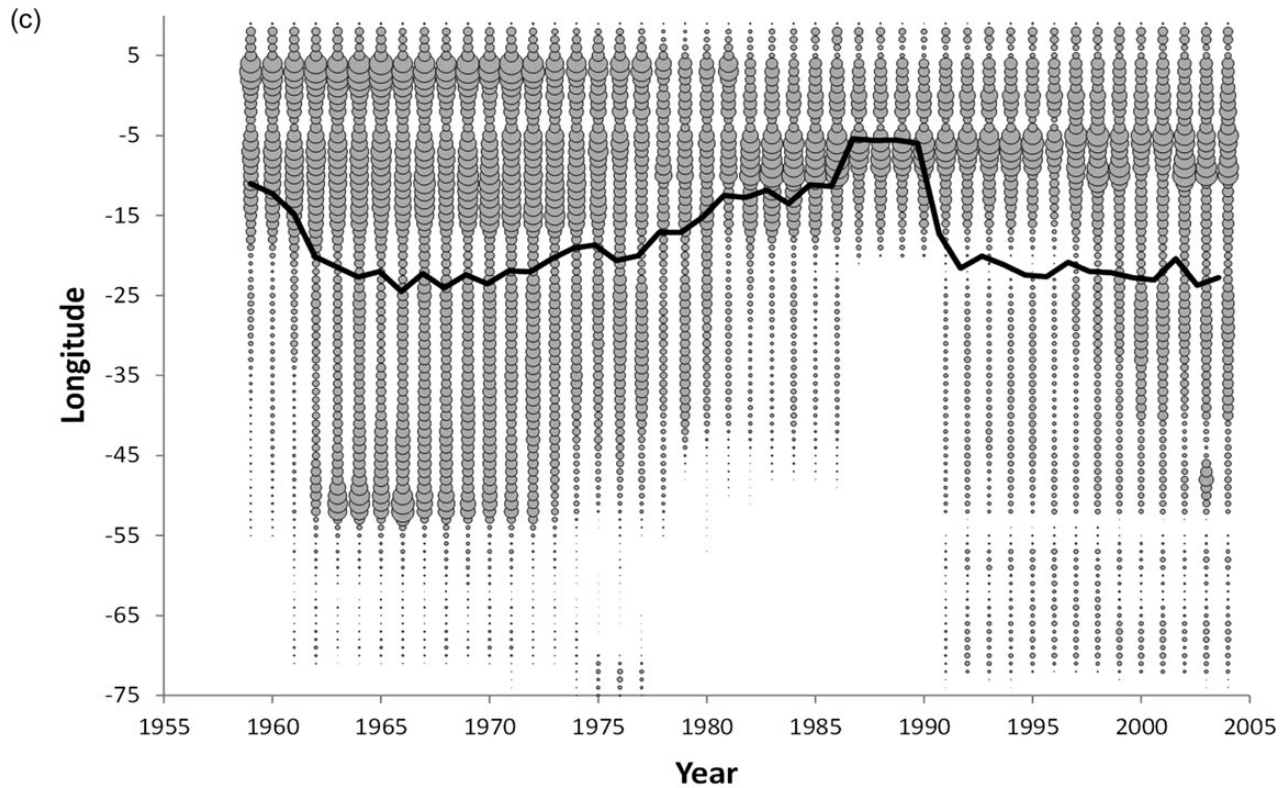


Figure 3. (Continued).

Table 1. Explained deviance of *Calanus* spp. occurrence by each factor.

	<i>C. glacialis</i>	<i>C. hyperboreus</i>	<i>C. finmarchicus</i>	<i>C. helgolandicus</i>
PCI	0.27	0.16	0.16	4.15
PCI + diatoms + dinoflagellates	0.51	1.41	2.41	6.41
SST	26.3	17.7	11.6	10.0
Salinity	19.4	10.7	11.4	2.05
Bathymetry	1.06	3.59	4.72	15.7
Vertical velocity	0.53	1.32	1.09	5.47
Spatial terms (Lat:Long)	35.6	23.1	26.3	39.0
Month	3.78	19.3	3.91	0.55
Year	4.87	0.75	1.66	2.74

model (adding 0.9–2.2% to the explained variance), indicating that high covariation exists between both variables and types and that spatial and temporal terms are key to reconstruct habitat suitability over the period. For the combined model, the estimated degrees of freedom, p -value for each selected variables, and the accuracy measures of cross-validation are indicated in Table 2. According to the AUC and overall accuracy, models for *C. finmarchicus*, *C. glacialis*, and *C. helgolandicus* have good agreement between occurrence predictions and observations (AUC: 0.77–0.88 and overall accuracy: 78–79%), whereas that for *C. hyperboreus* was slightly less accurate (AUC: 0.72, overall accuracy: 72%). The high accuracy values for cross-validated models, also compared with that of model using all observations (80–90%), indicate that models do not present overfitting. As an example, Figure 4 shows the GAM response curves of the combined model for *C. helgolandicus* occurrence probability, whereas Figure 5 shows the SST model compared with the spatial and temporal model for the same species in June, averaged by periods. In particular from the SST model, it appears that the habitat suitability of

C. helgolandicus at the southern limit nearby the Bay of Biscay shifted poleward from the 1970s to the 2000s, whereas the spatial and temporal model indicates a general expansion of the population in all directions, southward, northward, and offshore.

Seasonal analysis

To assess if the gravity centre index used is a good descriptor of the shift, first, the indices were applied to the *Calanus* spp. populations from the habitat suitability models (combined model) for each time-step (i.e. a month) then represented along the seasonal cycle (Figure 6). Population gravity centre shifts poleward from April to August (average for the overall 1959–2004 period) and southward from September to February. This pattern is consistent in all four species, with low differences in the magnitude of this shift (i.e. from 2° of latitude range in *C. finmarchicus* to 4° in *C. helgolandicus* and *C. hyperboreus*). This shift corresponds to the population growth differential over the latitude and not an advection of individuals. For *C. finmarchicus*, *C. glacialis*, and *C. hyperboreus*, it also

Table 2. Combined GAMs.

Species	Variables selected	EDF	p-value	Overall explained deviance (%)	AUC (model with all observations/mean kfold cross-validation)	Accuracy (model with all observations/mean kfold cross-validation) (%)
<i>C. glacialis</i>	Year	3.979	< 1.10 ⁻¹⁵	42.6	0.888/0.774	89.1/78.2
	SST	1.361	< 1.10 ⁻¹⁵			
	Salinity	1.289	0.0019			
	Longitude	2.964	< 1.10 ⁻¹⁵			
	Month		10 months significant			
<i>C. finmarchicus</i>	Year	3.940	< 1.10 ⁻¹⁵	34.4	0.778/0.778	79.8/79.8
	SST	1.997	< 1.10 ⁻¹⁵			
	Long:Lat	14.957	< 1.10 ⁻¹⁵			
	Month		10 months significant			
<i>C. hyperboreus</i>	Year	2.759	< 1.10 ⁻¹⁵	46.0	0.889/0.717	90.5/72.2
	SST	1.979	< 1.10 ⁻¹⁵			
	Long:Lat	14.771	< 1.10 ⁻¹⁵			
	Month		9 months significant			
<i>C. helgolandicus</i>	Year	3.988	< 1.10 ⁻¹⁵	41.2	0.819/0.798	80.8/79.7
	SST	1.998	< 1.10 ⁻¹⁵			
	Long:Lat	14.950	< 1.10 ⁻¹⁵			

Variables entered: Year, month, longitude, latitude, SST, salinity, bathymetry, vertical velocity. EDF, Estimated Degrees of Freedom. Thresholds for the conversion of probability of species presence to either the presence or the absence in model validation: 0.05 (*C. glacialis*), 0.60 (*C. finmarchicus*), 0.03 (*C. hyperboreus*), and 0.30 (*C. helgolandicus*).

may correspond to the ascent of the overwintering population at the end of the diapause (Heath *et al.*, 2004) rather than to a shift in the distribution of individuals remaining in surface during winter (the CPR only samples the surface layer). The gravity centre index used for *Calanus* spp. showed seasonal patterns which are consistent with the annual peaks reported for those taxa at different latitudes by other authors (Planque and Fromentin, 1996).

Temporal analysis

Population gravity centre indices throughout the years (Figure 7, Table 3) permitted to infer the following trends for each species according to the comparison between models and observed data. For *C. glacialis*, all models and observed (presence/absence) data showed a slight shift to the South (Figure 7a and e). This species presented relatively low occurrence in the data (i.e. *C. glacialis* was present in 5023 samples within the entire spatial domain and time-series, absent during 9 years) compared with other species (e.g. *C. finmarchicus* recorded in 105 598 samples in all years). This species has its main geographic distribution at Arctic and Subarctic regions; hence, enlarging the domain area of study beyond 65° is needed to confirm this result. For *C. hyperboreus*, observed data and the model based on the spatial and temporal terms indicate a shift to the South (Figure 7b and f). The combined model and the SST model showed no significant shift (Figure 7b). This means that species distribution probably shifted slightly to the Equator, while its habitat suitability was not expected to shift. This species, which presented relatively low occurrence in the data (present in 3853 samples, absent during 4 years), has its main geographic distribution in the Arctic and Subarctic regions; hence, enlarging the domain area of study beyond 65° is needed to confirm this result. For *C. finmarchicus*, all models showed poleward shift (Figure 7c and g). This means that species populations and its habitat suitability shifted poleward; hence, we can infer that population shift can be associated

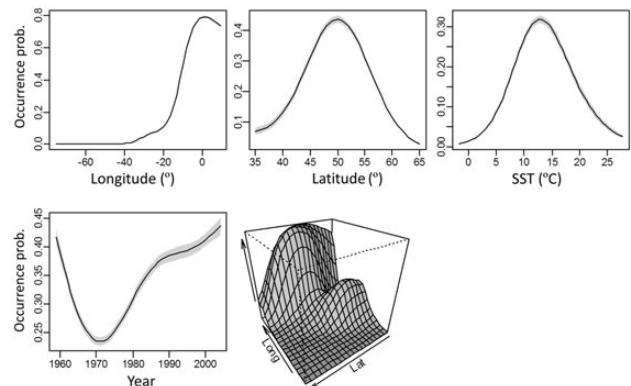


Figure 4. Response of *C. helgolandicus* occurrence probability for each variable of the combined model.

with sea warming. In particular, the spatial and temporal model indicates a population shift rate of 0.0073° latitude per year (8.1 km per decade) for the overall North Atlantic, and a slightly higher for the Northeast Atlantic (longitude between 30°W and 15°E; 16.5 km per decade) (Figure 7c). On the contrary, observed data showed that the gravity centre has a tendency to shift to the South (Figure 7g). This southward shift is due to the latitudinal decreasing trend in the sampling frequency (Figure 3b, Supplementary SI2), particularly north of 64° Latitude in the Northeastern Atlantic, in recent years, see Supplementary SI2. For *C. helgolandicus*, habitat suitability based on the SST model showed poleward shift since the 1970s; the combined model showed a similar result but the shift rate is less and starts from the 1980s (Figure 7d). On the contrary, the model based on spatial and temporal terms did not show a poleward shift. Observed data of *C. helgolandicus* showed slight southward shift for the presence/absence data, and no shift considering abundance data (Figure 7h).

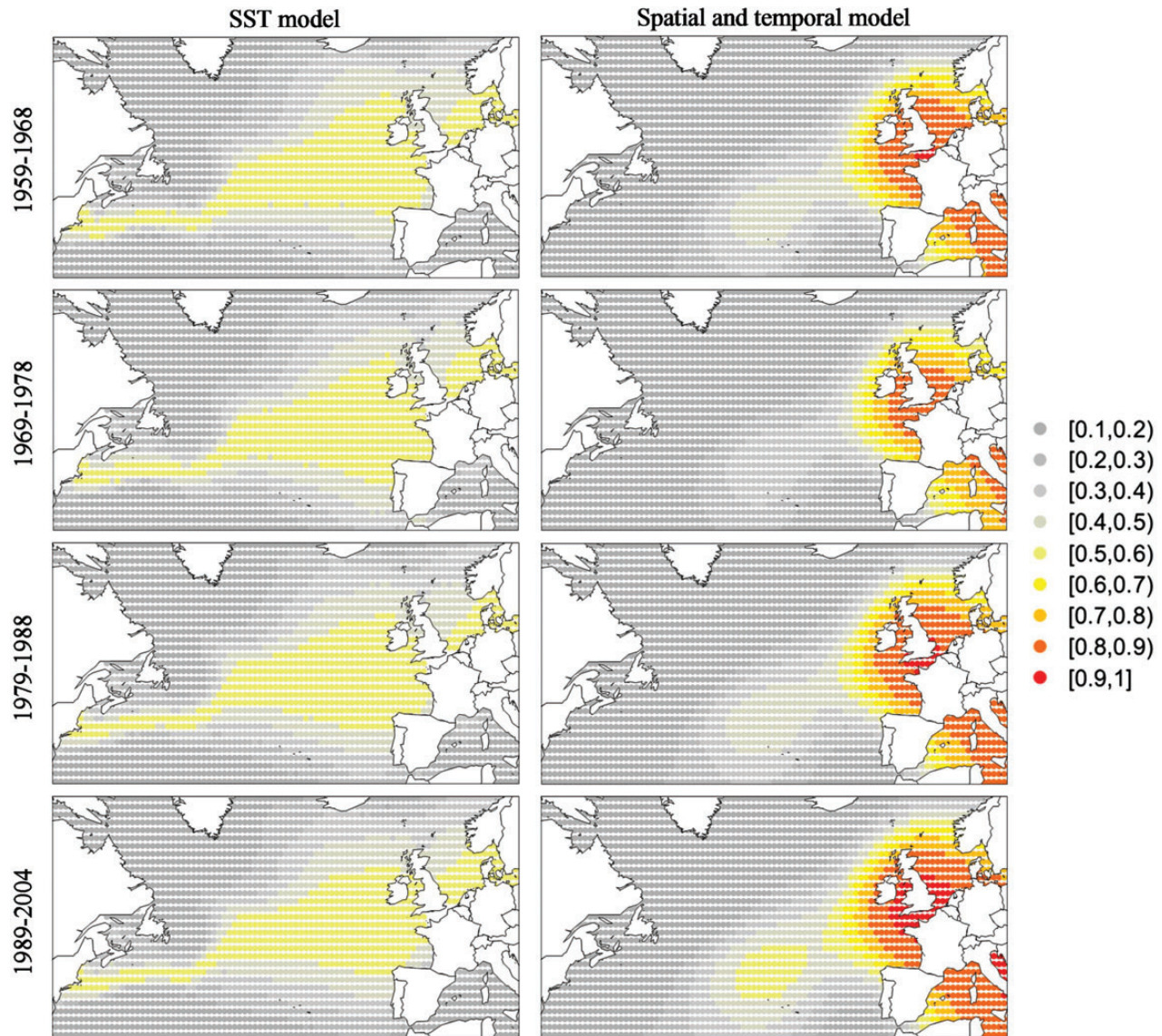


Figure 5. Evolution of *C. helgolandicus* occurrence probability models (SST model, as habitat suitability model, and spatial and temporal model as population data reconstruction) between 1959 and 2004 (month: June).

Overall, this means that species population probably did not shift poleward, and only its habitat suitability is shifted by sea warming.

Discussion and conclusion

GAMs for the four *Calanus* spp. in the North Atlantic have been built to analyse species distribution shifts between 1959 and 2004. GAMs were previously used to investigate the influence of environmental factors on plankton standing stock in the North Sea (Llope *et al.*, 2009, 2012). Concerning the species habitat suitability, the limitation of distance-based habitat models used in previous works (e.g. Helaouet and Beaugrand, 2007, 2009; Beaugrand *et al.*, 2008; Helaouet *et al.*, 2011; Reygondeau and Beaugrand, 2011) to describe the asymmetry of ecological niche has been well solved by GAMs, which are usually more accurate according to model comparison studies (e.g. Elith *et al.*, 2006). The models combining environmental and spatial factors have been validated using independent datasets and their reliability measure indicate moderate (72% for *C. hyperboricus*) to high accuracy performance (78–79% for *C. finmarchicus*,

C. glacialis, and *C. helgolandicus*). The reliability of our data reconstruction models and the gravity centres used here, encompassing the overall population in the study area, might improve previous estimates of shift rates based only on observed data (e.g. Beaugrand *et al.*, 2002).

Among the four analysed species, only *C. finmarchicus* showed a consistent poleward shift during the 1959–2004 period, as shown by all three models. This suggests that *C. finmarchicus* populations shifted poleward as a consequence of sea warming, since the SST model (i.e. habitat suitability) is consistent with the model with spatial and temporal terms (i.e. population data reconstruction; see maps in Supplementary SI3). This poleward shift is, overall, in agreement with previous works (see a review in Richardson, 2008). However, the rate of northward movement detected in *C. finmarchicus* (8.1 km per decade for the overall North Atlantic, 16.5 per decade for the Northeast Atlantic) is substantially lower than the change in distribution indicated for zooplankton assemblages by previous works for the Northeast Atlantic (260 km per decade for

the northward extension of zooplankton assemblages 14 species including the four *Calanus* spp. analysed here; [Beaugrand et al., 2002](#)). The main differences between those studies are the taxa assemblage analysed, the statistic considered (distribution limits vs. geographic centres), and the model algorithm used. The gravity centre

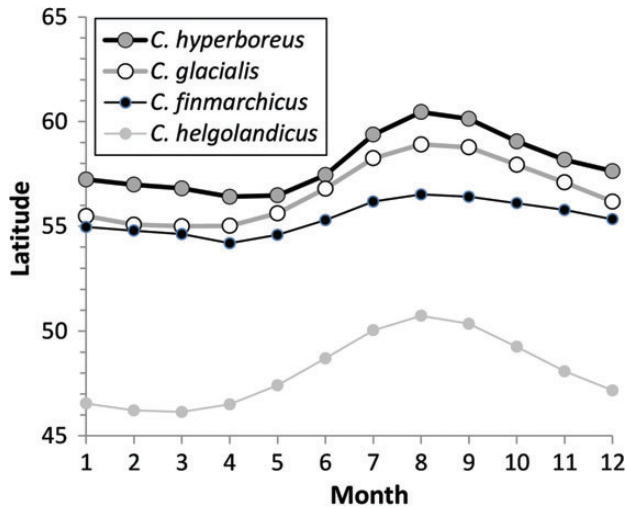


Figure 6. Latitudinal shift of the population gravity centre during seasonal cycle (average for the overall 1959–2004 period; combined model). See text for explanation.

considered here is more reliable to capture overall population movement, since distribution limits may also change because overall increase in population abundance. The rate estimate here for *C. finmarchicus* is still higher than that reported for terrestrial species based on a study of 99 species of birds, butterflies, and plants (6.1 km per decade, [Parmesan and Yohe, 2003](#)) but within that of rattlesnakes (4.3–24.2 km per decade, [Lawing and Polly, 2011](#)). On the contrary, the population centre gravity in the observed data of *C. finmarchicus* showed a southward shift. This is most likely due to the lower sampling frequency at latitudes >64°N after the early-1980s (Figure 3b), as shown by the long-term mean latitude of sampling frequency (Supplementary SI2a and b); indeed, the Norwegian Sea, which is a core area of distribution for *C. finmarchicus* in the Northeast Atlantic, was poorly sampled between 1981 and 2004. Therefore, our data reconstruction modelling approach, which is supported by high model accuracy, shows its power in detecting latitudinal shifts even when observations are lacking due to discontinuous sampling.

Habitat suitability of *C. helgolandicus* based on SST showed poleward shift since 1970s. On the contrary, the model based on spatial and temporal terms did not show poleward shift, but an expansion of the overall population in the North Sea, Atlantic margin, and central N Atlantic since the 1970s (as shown by the predicted maps in Figure 7d and by the increasing variance of the gravity centre not shown here for brevity). This means that *C. helgolandicus* population did not shift poleward, although its habitat suitability changed probably as a result of sea warming. The causes of

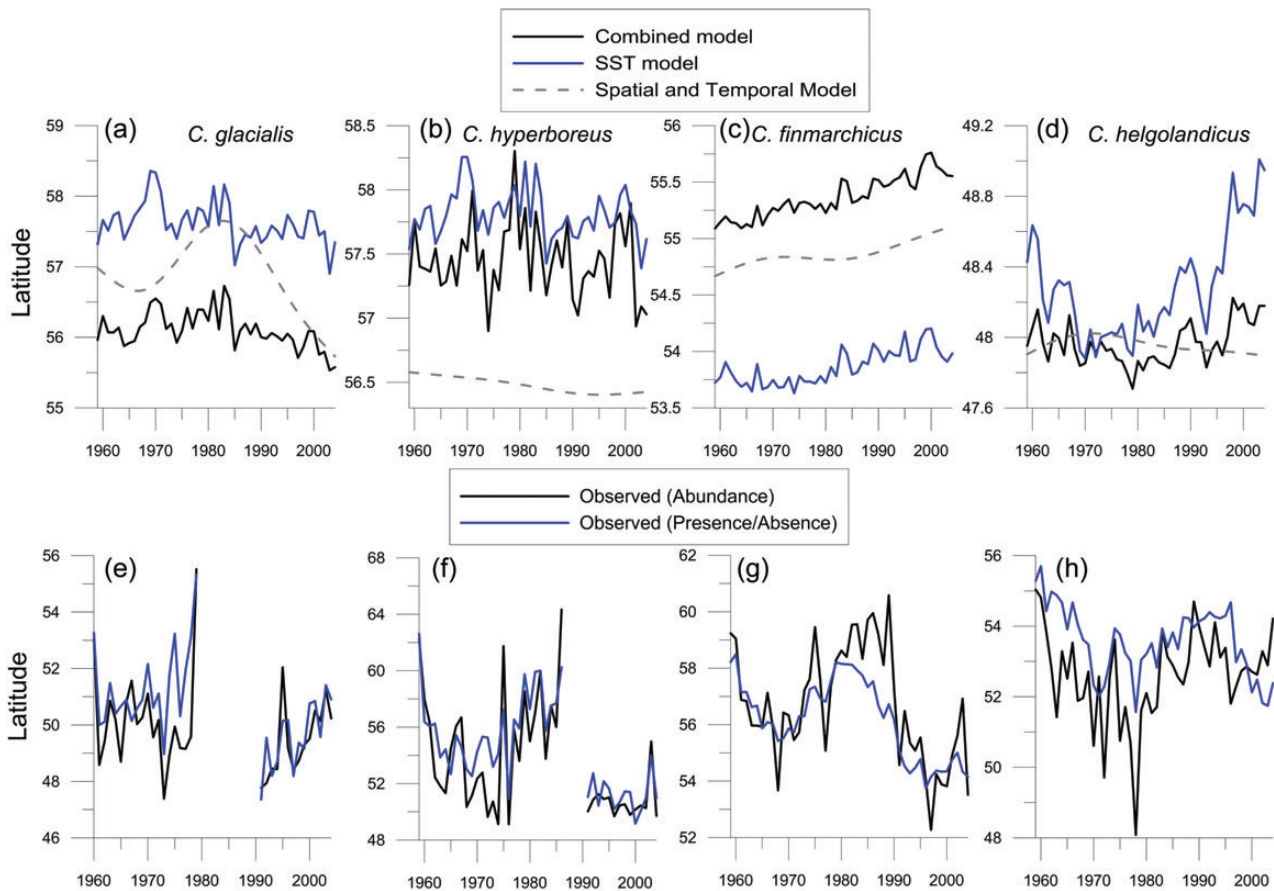


Figure 7. Shift in latitude of the gravity centre of *Calanus* spp. population habitat suitability models (a–d) and of the observed annual average (e–h), according to abundance and the presence/absence data. See estimations of trends (and their statistical significance) in [Table 3](#).

Table 3. Latitudinal shifts ($^{\circ}$ year $^{-1}$) for each species and according to observed data and models between 1959 and 2004 in the North Atlantic.

	<i>C. glacialis</i>		<i>C. hyperboreus</i>		<i>C. finmarchicus</i>		<i>C. helgolandicus</i>	
	Trend	p-value	Trend	p-value	Trend	p-value	Trend	p-value
Observed data (Pres/Abs)	-0.0386	0.0369	-0.1194	0.0006	-0.0641	4.5e-06	-0.0304	0.0042
Observed data (abundance)	-0.0174	0.3570	-0.0979	0.0252	-0.0483	0.0296	0.0119	0.4170
SST model	-0.0077	0.0178	-0.0024	0.2770	0.0086	1.9e-09	0.0119	0.0002
Spatial and temporal model	-0.0143	0.0136	-0.0044	<2e-16	0.0073	<2e-16	-0.0019	1.2e-06
Combined model	-0.0079	0.0049	-0.0020	0.5350	0.0123	<2e-16	0.0029	0.0261

C. helgolandicus population expansion in all directions, including southward, are still unclear; one could hypothesize that it could be partially related to the general increase in phytoplankton biomass observed since the mid-1980s in the Northeast Atlantic (Raitos et al., 2005; McQuatters-Gollop et al., 2011). This is supported by the fact that the CPR phytoplankton indices explained *C. helgolandicus* occurrence in a larger extent than for the other three species (Table 1). The expansion of *C. helgolandicus* includes its northern distribution limit in the North Sea where it has replaced *C. finmarchicus* since the 1990s, which has been interpreted as a result of warming of this region (Richardson, 2008). Changes in abundance are more difficult to attribute to global warming than are shifts in distribution or phenology (Richardson, 2008). Our approach helps to disentangle shifts in species distribution and shifts in their habitats, reconstructing species expected occurrence even when observations are not available.

Southward population movement of the Arctic species (*C. glacialis* and *C. hyperboreus*) that have their main geographic centre at NW Atlantic is in agreement with previous works focused on the NW Atlantic (see Arctic and Subarctic assemblages in Beaugrand et al., 2002, and *C. hyperboreus* in Johns et al., 2001). The southward shift of these two species would respond to cool water penetrating southward in the NW Atlantic (Richardson, 2008), in particular in the Scotian and Newfoundland shelf regions influenced by the outflow of freshwater from the Arctic (Head and Pepin, 2010; Licandro et al., 2011). The spatial difference of SST between cold (1969–1979) and warm (1998–2004) periods (Figure 2b) showed that warming in the NW Atlantic (Labrador Sea) was more heterogeneous than in the NE Atlantic with three hotspots of warming and a region of cooling. This pattern could be related to the observed trend of the North Atlantic Oscillation (NAO) index towards positive values (Beaugrand et al., 2002) that result in warmer winters in the NE Atlantic and colder winters in the NW Atlantic (Hurrell, 1995) and might explain why *C. glacialis* and *C. hyperboreus* are showing southward shift. On the other hand, these two Arctic species were less frequent in the time-series (9 and 4 years without any presence for *C. glacialis* and *C. hyperboreus*, respectively) as their main geographic distribution is usually in Arctic and Subarctic regions, out of the area sampled by the CPR; hence, enlarging the domain area of study beyond 65° is needed to confirm these results.

Latitudinal shift of species is not the only potential response to ocean warming. Marine fish and invertebrates, for instance, can also respond to warming migrating to deeper waters (Cheung et al., 2013), similar to terrestrial species shifting their distribution to higher elevations within mountains (Engler et al., 2011). Using electronic tagging, Neat and Righton (2007) found that cod moved to cooler water at greater depths. The limitations of our data based uniquely on surface sampling (Jónasdóttir and Koski,

2011) preclude analysing whether there is also a shift towards deeper waters. In particular, Williams (1985) and Jónasdóttir and Koski (2011) observed that when *C. helgolandicus* and *C. finmarchicus* co-occur, the latter preferred colder deeper waters remaining reproductively active, while the former stayed in the warmer surface waters. In *C. glacialis*, Niehoff and Hirche (2005) found that temperature increase in the surface layer apparently triggered the descent of the females to lower depths and the arrest of their reproductive activity. In *Neocalanus plumchrus*, the timing of the annual maximum peak biomass has shifted 60 d earlier in warm than in cold years over the past 50 years (Mackas et al., 1998). Such phenological change is probably a consequence of both increased survivorship of early cohorts and increase in physiological rates such as egg hatching, reproductive, and growth rate (Richardson, 2008). According to the study by Helouet and Beaugrand (2009), the application of the ecological niche theory predicts that for northern hemisphere species, warming should produce increased survivorship in the northern (i.e. at cooler waters) population side in respect with its optimal niche, while it will diminish survivorship at the southern (i.e. at warmer waters) population side. Similarly, for competing species warming should produce increased competitive advantage for southern species with respect to northern species in overlapping areas. These two processes should cause latitudinal shift of the overall species distribution, thus, keeping pace with the shift of their suitable habitats. If vertical migration plays also a role in the climate response, even assuming that *Calanus* spp. could remain deeper ingesting an adequate ration of food, a drop in overall occurrence frequency at the two (northern-southern) sides of optimal (realized) niche should be observed in CPR surface-based data, which is not the case for *C. helgolandicus* according to our results.

Among the environmental factors, SST is, in general, the one explaining most of the variance of species occurrence in the four *Calanus* spp. (especially in *C. glacialis*), in agreement with previous works (Beaugrand et al., 2013) and the view of temperature as the most broadly influential factor controlling biological processes (Brown et al., 2004). Other environmental factors that could be taken into account in future studies to improve trend analysis and habitat suitability models are horizontal currents, food resources, and suitability of the overwintering habitat (Irigoién, 2004). As autonomous motility of copepods is spatially limited when compared with passive movement triggered by ocean currents, dispersal by this means might influence the probability of occurrence of a species at a given location from month to month. Phytoplankton indices used here and *Calanus* occurrence were poorly related according to our modelling approach as in previous attempts (e.g. Reygondeau and Beaugrand, 2011), either because those indices represent only part of the food available or because food is not a limiting factor. Phytoplankton is a key food source for calanoid species (Moller

et al., 2012); hence, other phytoplankton indices (satellite-derived primary production) or other ways of modelling food resources in the habitat suitability model should be explored in the future. Further, factors affecting overwintering survival such as winter mixing depth or distribution of predators might also play a role in the final distribution (Irigoien, 2004).

Another relevant finding of this study is that environmental variables accounted for 17–31% of species occurrence, which is less than did spatial and temporal terms (33–45%) in all *Calanus* spp., with high covariation among both types of factors. This means that space and seasonality are key to describe the pelagic distribution of copepods. Also, since spatial and temporal terms can be partially a proxy of dispersal constraints, our results are in line with a balanced view between the importance of the role of dispersal limitation and of niche partitioning on the plankton community structure (see for instance, Irigoien et al., 2011, for zooplankton and Chust et al., 2013, for phytoplankton).

Since species habitat models developed here have high accuracy, according to model evaluation using independent datasets, they have also the potential to be applied to future climate change scenarios to assess warming impacts on zooplankton and to assess the implications of changes at this trophic level vs. top predators such as fish and cetaceans. As we have seen, warming can modify the distribution of marine organisms, which in turn, it can be propagated through the upper trophic levels. In a context of rapid alteration of marine ecosystems throughout the world (Pauly et al., 1998), future projections of ocean productivity, based on habitat species distribution, are needed for a detailed assessment of ocean health and benefits and for achieving or maintaining the good environmental status of the North Atlantic (see for instance, the environmental status defined by the Marine Strategy Framework Directive, MSFD; European Commission, 2008).

In summary, our findings show that among the four *Calanus* spp. analysed in the North Atlantic during the 1959–2004 period, only *C. finmarchicus* had a consistent poleward shift estimated in 8.1 km per decade (16.5 km per decade for the Northeast Atlantic), triggered by sea warming, which is lower than previous works restricted to the Northeast Atlantic and considering zooplankton assemblages. Because of the sampling limitation of the CPR survey in the northern regions, it is not possible to clarify whether *C. finmarchicus* global gravity centre has displaced to the North or rather its southern distribution has retracted. Our model is highly accurate (80%), which has been validated by a subset of observational data. It shows its power in detecting latitudinal shifts even when changes in the sampling frequency may not ensure an adequate coverage of a region. On the contrary, *C. helgolandicus* population gravity centre did not shift poleward according to the results, although its northern distribution limit in the North Sea has shifted northward. *Calanus helgolandicus* is expanding in all directions probably influenced by multiple factors (e.g. phytoplankton increase, warming, their own population dynamics, biotic interactions, environmental variability). *Calanus glacialis* and *C. hyperboreus* based on the CPR sampling, present the geographic centres of their populations mainly in the NW Atlantic, showed a slight southward shift, probably responding to cool water penetrating south in the Labrador Current and the heterogeneity in the warming at NW Atlantic. Despite the limitations of CPR data restricted to surface sampling, the long and spatially extensive biological dataset, together with species occurrence models based on a wide range of factors (climate, spatio-temporal) and tested in a combined and separate

manner, has the potential to detect and accurately quantify latitudinal shifts and suggests potential causes.

Supplementary data

Supplementary material is available at the *ICESJMS* online version of the manuscript.

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References

- Amante, C., and Eakins, B. W. 2009. ETOPO1 1 Arc-Minute Global Relief Model: Procedures, Data Sources and Analysis. NOAA Technical Memorandum, NESDIS NGDC-24. 19 pp.
- Batten, S. D., Clark, R., Flinkman, J., Hays, G. C., John, E., John, A. W. G., Jonas, T., et al. 2003. CPR sampling: the technical background, materials and methods, consistency and comparability. *Progress in Oceanography*, 58: 193–215.
- Beaugrand, G., Edwards, M., Brander, K., Luczak, C., and Ibanez, F. 2008. Causes and projections of abrupt climate-driven ecosystem shifts in the North Atlantic. *Ecology Letters*, 11: 1157–1168.
- Beaugrand, G., Mackas, D., and Goberville, E. 2013. Applying the concept of the ecological niche and a macroecological approach to understand how climate influences zooplankton: advantages, assumptions, limitations and requirements. *Progress in Oceanography*, 111: 75–90.
- Beaugrand, G., Reid, P. C., Ibáñez, F., Lindley, J. A., and Edwards, M. 2002. Reorganization of North Atlantic marine copepod biodiversity and climate. *Science*, 296: 1692–1694.
- Brown, J. H., Gillooly, J. F., Allen, A. P., Savage, V. M., and West, G. B. 2004. Toward a metabolic theory of ecology. *Ecology*, 85: 1771–1789.
- Burnham, K. P., and Anderson, D. R. 2002. *Model Selection and Multi-modal Inference: a Practical Information — Theoretic Approach*, 2nd edn. Springer, New York.
- Cheung, W. W. L., Lam, V. W. Y., Sarmiento, J. L., Kearney, K., Watson, R., and Pauly, D. 2009. Projecting global marine biodiversity impacts under climate change scenarios. *Fish and Fisheries*, 10: 235–251.
- Cheung, W. W. L., Watson, R., and Pauly, D. 2013. Signature of ocean warming in global fisheries catch. *Nature*, 497: 365–368.
- Chust, G., Irigoien, X., Chave, J., and Harris, R. P. 2013. Latitudinal phytoplankton distribution and the neutral theory of biodiversity. *Global Ecology and Biogeography*, 22: 531–543. doi:10.1111/geb.12016.
- Congalton, R. G. 1991. A review of assessing the accuracy of classifications of remotely sensed data. *Remote Sensing of Environment*, 37: 35–46.
- Elith, J., Graham, C. H., Anderson, R. P., Dudík, M., Ferrier, S., Guisan, A., Hijmans, R. J., et al. 2006. Novel methods improve prediction of species' distributions from occurrence data. *Ecography*, 29: 129–151.
- Elith, J., and Leathwick, J. R. 2009. Species distribution models: ecological explanation and prediction across space and time. *Annual Review of Ecology Evolution and Systematics*, 40: 677–697.
- Engler, R., Randin, C. F., Thuiller, W., Dullinger, S., Zimmermann, N. E., Araujo, M. B., Pearman, P. B., et al. 2011. 21st century climate change threatens mountain flora unequally across Europe. *Global Change Biology*, 17: 2330–2341.
- European Commission. 2008. Directive 2008/56/EC of the European Parliament and of the Council of 17 June 2008, establishing a framework for community action in the field of marine environmental

- policy (Marine Strategy Framework Directive). Official Journal of the European Union, L164: 19–40.
- Fielding, A. H., and Bell, J. F. 1997. A review of methods for the assessment of prediction errors in conservation presence/absence models. *Environmental Conservation*, 24: 38–49.
- Guisan, A., and Zimmermann, N. E. 2000. Predictive habitat distribution models in ecology. *Ecological Modelling*, 135: 147–186.
- Hastie, T. J., and Tibshirani, R. J. 1990. *Generalized Additive Models*. Chapman and Hall, London.
- Head, E. J. H., Harris, L. R., and Campbell, R. W. 2000. Investigations on the ecology of *Calanus* spp. in the Labrador Sea. I. Relationship between the phytoplankton bloom and reproduction and development of *Calanus finmarchicus* in spring. *Marine Ecology Progress Series*, 193: 53–73.
- Head, E. J. H., and Pepin, P. 2010. Spatial and inter-decadal variability in plankton abundance and composition in the Northwest Atlantic (1958–2006). *Journal of Plankton Research*, 32: 1633–1648.
- Heath, M. R., Boyle, P. R., Gislason, A., Gurney, W. S. C., Hay, S. J., Head, E. J. H., Holmes, S., et al. 2004. Comparative ecology of overwintering *Calanus finmarchicus* in the northern North Atlantic, and implications for life-cycle patterns. *ICES Journal of Marine Science*, 61: 698–708.
- Helaouet, P., and Beaugrand, G. 2007. Macroecology of *Calanus finmarchicus* and *C-helgolandicus* in the North Atlantic Ocean and adjacent seas. *Marine Ecology Progress Series*, 345: 147–165.
- Helaouet, P., and Beaugrand, G. 2009. Physiology, ecological niches and species distribution. *Ecosystems*, 12: 1235–1245.
- Helaouet, P., Beaugrand, G., and Reid, P. C. 2011. Macrophysiology of *Calanus finmarchicus* in the North Atlantic Ocean. *Progress in Oceanography*, 91: 217–228.
- Hijmans, R. J., Phillips, S., Leathwick, J., and Elith, J. 2012. Dismo: species distribution modelling. R package version 0.7-17. <http://CRAN.R-project.org/package=dismo>.
- Hirzel, A. H., Hausser, J., Chessel, D., and Perrin, N. 2002. Ecological-niche factor analysis: how to compute habitat-suitability maps without absence data? *Ecology*, 83: 2027–2036.
- Hurrell, J. W. 1995. Decadal trends in the North Atlantic Oscillation: regional temperatures and precipitation. *Science-AAAS-Weekly Paper Edition*, 269.5224: 676–678.
- Hutchinson, G. E. 1957. Concluding remarks. *Cold Spring Harbor Symposia on Quantitative Biology*, 22: 415–427.
- Irigoien, X. 2004. Some ideas about the role of lipids in the life cycle of *Calanus finmarchicus*. *Journal of Plankton Research*, 26: 259–263.
- Irigoien, X., Chust, G., Fernandes, J. A., Albaina, A., and Zarauz, L. 2011. Factors determining mesozooplankton species distribution and community structure in shelf and coastal waters. *Journal of Plankton Research*, 33: 1182–1192.
- Irigoien, X., Head, R., Klenke, U., Meyer-Harms, B., Harbour, D., Niehoff, B., Hirche, H. J., et al. 1998. A high frequency time series at weather ship M, Norwegian Sea, during the 1997 spring bloom: feeding of adult female *Calanus finmarchicus*. *Marine Ecology Progress Series*, 172: 127–137.
- Isaaks, E. H., and Srivastava, R. M. 1989. *An Introduction to Applied Geostatistics*. Ed. by E. H. Isaaks, and R. Mohan Srivastava. Oxford University Press, Oxford, UK. 561 pp.
- Jimenez-Valverde, A., and Lobo, J. M. 2007. Threshold criteria for conversion of probability of species presence to either-or presence-absence. *Acta Oecologica*, 31: 361–369.
- Johns, D. G., Edwards, M., and Batten, S. D. 2001. Arctic boreal plankton species in the Northwest Atlantic. *Canadian Journal of Fisheries and Aquatic Sciences*, 58: 2121–2124.
- Jonas, T. D., Walne, A., Beaugrand, G., Gregory, L., and Hays, G. C. 2004. The volume filtered by a continuous plankton recorder sample: the effect of ship speed. *Journal of Plankton Research*, 26: 1499–1506.
- Jónasdóttir, S. H., and Koski, M. 2011. Biological processes in the North Sea: comparison of *Calanus helgolandicus* and *Calanus finmarchicus* vertical distribution and production. *Journal of Plankton Research*, 33: 85–103.
- Lawing, A. M., and Polly, P. D. 2011. Pleistocene climate, phylogeny, and climate envelope models: an integrative approach to better understand species' response to climate change. *PLoS One*, 6: e28554.
- Legendre, P. 1993. Spatial autocorrelation: trouble or new paradigm? *Ecology*, 74: 1659–1673.
- Lewis, K., Allen, J. I., Richardson, A. J., and Holt, J. T. 2006. Error quantification of a high resolution coupled hydrodynamic-ecosystem coastal-ocean model: Part 3, validation with Continuous Plankton Recorder data. *Journal of Marine Systems*, 63: 209–224.
- Licandro, P., Head, E., Gislason, A., Benfield, M. C., Harvey, M., Margonski, P., and Silke, J. 2011. Overview of trends in plankton communities. In *ICES Status Report on Climate Change in the North Atlantic*, pp. 103–122. Ed. by P. C. Reid, and L. Valdés. ICES Cooperative Research Report, 310.
- Lindley, J. A., and Daykin, S. 2005. Variations in the distributions of *Centropages chierchiae* and *Temora stylifera* (Copepoda: Calanoida) in the north-eastern Atlantic Ocean and western European shelf waters. *ICES Journal of Marine Science*, 62: 869–877.
- Llope, M., Chan, K.-S., Ciannelli, L., Reid, P. C., Stige, L. C., and Stenseth, N. C. 2009. Effects of environmental conditions on the seasonal distribution of phytoplankton biomass in the North Sea. *Limnology and Oceanography*, 54: 512–524.
- Llope, M., Licandro, P., Chan, K.-S., and Stenseth, N. C. 2012. Spatial variability of the plankton trophic interaction in the North Sea: a new feature after the early 1970s. *Global Change Biology*, 18: 106–117.
- Mackas, D. L., Goldblatt, R., and Lewis, A. G. 1998. Interdecadal variation in developmental timing of *Neocalanus plumchrus* populations at Ocean Station P in the Subarctic North Pacific. *Canadian Journal of Fisheries and Aquatic Sciences*, 55: 1878–1893.
- McQuatters-Gollop, A., Reid, P. C., Edwards, M., Burkill, P. H., Castellani, C., Batten, S., Gieskes, W., et al. 2011. Is there a decline in marine phytoplankton? *Nature*, 472: E6–E7.
- Møller, E. F., Maar, M., Jónasdóttir, S. H., Nielsen, T. G., and Tønnesson, K. 2012. The effect of changes in temperature and food on the development of *Calanus finmarchicus* and *Calanus helgolandicus* populations. *Limnol. Oceanogr*, 57: 211–220.
- Neat, F., and Righton, D. 2007. Warm water occupancy by North Sea cod. *Proceedings of the Royal Society of London, Series B: Biological Sciences*, 274: 789–798.
- Niehoff, B., and Hirche, H.-J. 2005. Reproduction of *Calanus glacialis* in the Lureford (western Norway): indication for temperature-induced female dormancy. *Marine Ecology Progress Series*, 285: 107–115.
- Ohman, M. D., and Hirche, H.-J. 2001. Density-dependent mortality in an oceanic copepod population. *Nature*, 412.6847: 638–641.
- Oksanen, J., and Minchin, P. R. 2002. Continuum theory revisited: what shape are species responses along ecological gradients? *Ecological Modelling*, 157: 119–129.
- Parmesan, C., and Yohe, G. 2003. A globally coherent fingerprint of climate impacts across natural systems. *Nature*, 421: 37–42.
- Pauly, D., Christensen, V., Dalsgaard, J., Froese, R., and Torres, F., Jr. 1998. Fishing down marine food webs. *Science*, 279: 860–863.
- Pearson, R. G., and Dawson, T. P. 2003. Predicting the impacts of climate change on the distribution of species: are bioclimate envelope models useful? *Global Ecology and Biogeography*, 12: 361–371.
- Planque, B., and Batten, S. D. 2000. *Calanus finmarchicus* in the North Atlantic: the year of *Calanus* in the context of interdecadal change. *ICES Journal of Marine Science*, 57: 1528–1535.
- Planque, B., and Fromentin, J. 1996. *Calanus* and environment in the eastern North Atlantic. I. Spatial and temporal patterns of *C. finmarchicus* and *C. helgolandicus*. *Marine Ecology Progress Series*, 134: 101–109.
- Planque, B., Loots, C., Petitgas, P., Lindstrom, U., and Vaz, S. 2011. Understanding what controls the spatial distribution of fish populations using a multi-model approach. *Fisheries Oceanography*, 20: 1–17.
- Raes, N., and ter Steege, H. 2007. A null-model for significance testing of presence-only species distribution models. *Ecography*, 30: 727–736.

- Raitsos, D. E., Reid, P. C., Lavender, S. J., Edwards, M., and Richardson, A. J. 2005. Extending the SeaWiFS chlorophyll data set back 50 years in the northeast Atlantic. *Geophysical Research Letters*, 32: L06603.
- Reygondeau, G., and Beaugrand, G. 2011. Future climate-driven shifts in distribution of *Calanus finmarchicus*. *Global Change Biology*, 17: 756–766.
- Richardson, A. J. 2008. In hot water: zooplankton and climate change. *ICES Journal of Marine Science*, 65: 279–295.
- Richardson, A. J., Walne, A. W., John, A. W. G., Jonas, T. D., Lindley, J. A., Sims, D. W., Stevens, D., *et al.* 2006. Using continuous plankton recorder data. *Progress in Oceanography*, 68: 27–74.
- Runge, J. A., Franks, P. J., Gentleman, W. C., Megrey, B. A., Rose, K. A., Werner, F. E., and Zakardjian, B. (Ed.) 2005. *Diagnosis and Prediction of Variability in Secondary Production and Fish Recruitment Processes: Developments in Physical-Biological Modeling*. Harvard University Press, Cambridge, Massachusetts.
- Speirs, D. C., Gurney, W. S. C., Holmes, S. J., Heath, M. R., Wood, S. N., Clarke, E. D., Harms, I. H., *et al.* 2004. Understanding demography in an advective environment: modelling *Calanus finmarchicus* in the Norwegian Sea. *Journal of Animal Ecology*, 73: 897–910.
- Torgersen, T., and Huse, G. 2005. Variability in retention of *Calanus finmarchicus* in the Nordic Seas. *ICES Journal of Marine Science*, 62: 1301–1309.
- Valle, M., Borja, Á., Chust, G., Galparsoro, I., and Garmendia, J. M. 2011. Modelling suitable estuarine habitats for *Zostera noltii*, using Ecological Niche Factor Analysis and Bathymetric LiDAR. *Estuarine, Coastal and Shelf Science*, 94: 144–154.
- VanDerWal, J., Falconi, L., Januchowski, S., Shoo, L., and Storlie, C. 2012. Species Distribution Modelling Tools: tools for processing data associated with species distribution modelling exercises. R Package ‘SDMTools’. <http://cran.r-project.org/web/packages/SDMTools/SDMTools.pdf>.
- Warner, A. J., and Hays, G. C. 1994. Sampling by the continuous plankton recorder survey. *Progress in Oceanography*, 34: 237–256.
- Williams, R. 1985. Vertical distribution of *Calanus finmarchicus* and *C. helgolandicus* in relation to the development of the seasonal thermocline in the Celtic Sea. *Marine Biology*, 86: 145–149.
- Wuillez, M., Rivoirard, J., and Petitgas, P. 2009. Notes on survey-based spatial indicators for monitoring fish populations. *Aquatic Living Resources*, 22: 155–164.

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